

Prepared in cooperation with the Minnesota Pollution Control Agency

Wastewater Indicators, Hormones, Sterols, Antibiotics, and Pharmaceuticals in Soil at an Agricultural Field Irrigated with Domestic Septage, Central Minnesota, September 2014



Scientific Investigations Report 2018–5100

U.S. Department of the Interior U.S. Geological Survey

Cover. Collecting soil from an agricultural field in central Minnesota, September 2014. Photograph by Sarah M. Elliott, U.S. Geological Survey.

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Conversion Factors

International System of Units to U.S. customary units

| Multiply | Ву | To obtain |
|-----------------|----------|-----------------------|
| | Length | |
| centimeter (cm) | 0.3937 | inch (in.) |
| meter (m) | 3.281 | foot (ft) |
| | Volume | |
| milliliter (mL) | 0.033814 | ounce, fluid (fl. oz) |
| liter (L) | 0.2642 | gallon (gal) |

Temperature in degrees Celsius (°C) may be converted to degrees Fahrenheit (°F) as

 $^{\circ}F = (1.8 \times ^{\circ}C) + 32.$

Supplemental Information

Septage application rates are given in liters per hectare (L/ha).

Concentrations of wastewater indicators, hormones, sterols, antibiotics, and pharmaceuticals are given in micrograms per kilogram (μ g/kg), or milligrams per kilogram (mg/kg).

Wastewater Indicators, Hormones, Sterols, Antibiotics, and Pharmaceuticals in Soil at an Agricultural Field Irrigated with Domestic Septage, Central Minnesota, September 2014

By Sarah M. Elliott,¹ Melinda L. Erickson,¹ Aliesha L. Krall,¹ and Byron A. Adams²

Abstract

Treated domestic septage can be used to irrigate agricultural fields as a disposal method or as a means to reuse water. Because traditional on-site treatment systems are not designed to remove wastewater indicators, hormones, sterols, antibiotics, and pharmaceuticals, land application of septage potentially results in soil contamination. Soils were collected and analyzed from four sites in a central Minnesota agricultural field irrigated with domestic septage. Soil samples were analyzed for 111 unique contaminants, including wastewater indicators, hormones, sterols, antibiotics, and pharmaceuticals. In total, 32 contaminants were detected in soil samples. Several wastewater indicators were detected in soil, including fragrances, alkylphenols, and flame-retardants, at concentrations ranging from 1 (2,6-dimethylnaphthalene at soil site 4) to 1,550 (β -sitosterol at soil site 1) micrograms per kilogram. Relative to the number of contaminants analyzed, steroid hormones had the most frequent detections in soil samples (33 percent), and androgens were more prevalent compared to estrogens (50 and 22 percent, respectively). Androgens and estrogens were detected at concentrations ranging from 0.21 (estrone at soil site 3) to 3.9 (dihydrotestosterone at soil site 1) micrograms per kilogram. Quantifiable concentrations of antibiotics and pharmaceuticals ranged from 1.4 (carbamazepine at soil site 1) to 540 (azithromycin at soil site 3) micrograms per kilogram. Two antibiotics, ciprofloxacin and ofloxacin, were detected at concentrations above the limit of quantification (greater than 1,000 micrograms per kilogram at soil sites 2 and 3). This pilot sampling indicates that soils may be a repository for some contaminants introduced to the environment through land application of domestic septage.

Introduction

Domestic septage is a potential source of contaminants to soil when used for irrigation. Even after treatment, domestic septage is known to contain various pharmaceuticals, hormones, alkylphenols, and fragrances (Carrara and others, 2008; Katz and others, 2009; Lapworth and others, 2012). The practice of irrigating with domestic septage is an efficient reuse of water, especially in arid regions. However, the fate, transport, and ecological effects of septage-associated contaminants after application to the land surface are not well understood. Contaminants may accumulate in soils resulting in alterations to the microbial community, uptake by crops, or transport to underlying aquifers.

Unlike wastewater treatment plants that typically discharge into surface waters as a point source, agricultural lands can act as nonpoint sources of contaminants to the environment through runoff and infiltration. Herbicides, insecticides, and fungicides may be introduced to the environment through direct application to crops, whereas hormones and antibiotics may be introduced through animal excretion directly onto the land surface or through application of manure on a broad scale. Antibiotics commonly detected in stream waters and sediments in agricultural settings include sulfonamides and tetracyclines (Arikan and others, 2008; Bartelt-Hunt and others, 2011); however, most studies mainly focus on establishments that land apply or store livestock manure and less on sources such as treated domestic septage used for irrigation.

Irrigation using treated septage may accumulate contaminants in the soil (Chen and others, 2011). The affinity of organic chemicals to bind to soils depends on many factors including the physical properties of the chemical, soil carbon content, soil texture, and, in agricultural settings, field practices (Rodvang and Simpkins, 2001; Xu and others, 2009). For example, carbamazepine has a high affinity to sorb to soil with a high organic matter content, which often is present in the topmost layers of soil (Arye and others, 2011). However, depending on the type of tillage practices used, the

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organic-rich layer can be incorporated into deeper depths, affecting the chemical profile of the soil and environmental risks.

In addition to the potential ecological risks associated with applying domestic septage to land, potential humanhealth risks exist, such as contamination of downgradient drinking water and exposure through consumption (in agricultural settings). More than 15 million households in the United States rely on groundwater from private wells for domestic water use (U.S. Environmental Protection Agency, 2017); some of these wells are located downgradient and within the same aquifers as sites where domestic septage is applied to land. Pharmaceuticals, perfluorinated chemicals, and flameretardants represent some of the types of chemicals that may contaminate drinking water sources (Schaider and others, 2016); furthermore, domestic wells near a septic system and in shallow, thin aguifers are more vulnerable compared to others (Verstraeten and others, 2005). Another potential humanhealth risk is associated with consumption of produce grown on lands irrigated with domestic septage. Specifically, carbamazepine, carbamazepine metabolites, and lamotrigine can accumulate in certain vegetables (Malchi and others, 2014; Paz and others, 2016). As further evidence, carbamazepine and its metabolites were excreted from humans that consumed produce grown on lands irrigated with reclaimed wastewater (Paltiel and others, 2016).

The quantity of septage removed from septic tanks or applied to land in Minnesota is not tracked. However, based on the current estimate of 542,000 septic systems in the State, an approximate volume of septage pumped from tanks and other sanitation devices may be as high as 275 million gallons per year (Jensen, 2015). More than 400 licensed businesses (maintainers) pump and apply septage from septic tanks, holding tanks, and other sanitation devices from domestic and nondomestic sources in the State to land. Some maintainers apply septage to land at one site, whereas others may have many sites within their service areas, bringing the possible number of land application sites scattered across the State to more than 1,000 (Minnesota Pollution Control Agency, 2006). Given the estimated volume of septage that is applied to land in Minnesota, it is important to understand potential effects to the environment. A pilot study was completed to characterize the presence of wastewater indicators, hormones, sterols, antibiotics, and pharmaceuticals in soil collected from a central Minnesota agricultural field that irrigates with treated domestic septage.

Purpose and Scope

The purpose of this report is to present an assessment on the presence of wastewater indicators, hormones, sterols, antibiotics, and pharmaceuticals in soil collected from a Minnesota agricultural field that irrigates with treated domestic septage. The assessment was based on analyses of 111 unique contaminants in four soil samples collected in September 2014. Wastewater indicator, hormone, and sterol data are available in the USGS National Water Information System and can be searched by the following station numbers: 451700093430001 (soil site 1), 451700093440001 (soil site 2), 451700093430003 (soil site 3), 451700093430002 (soil site 4; U.S. Geological Survey, 2017). Antibiotic and pharmaceutical data are provided in this report.

Study Area

The study area is a 14-hectare agricultural field in central Minnesota, northwest of the Twin Cities Metropolitan Area (fig. 1). The field contains gently rolling, glacially deposited materials that extend to a depth of nearly 30 meters. Geologic materials identified during drilling of the four on-site monitoring wells included intermittent layers of sand and clays, sandy clays and sand, and gravel; no confining layers were identified. The depth to the water table at time of well installation was 12 to 15 meters below land surface (Minnesota Department of Health, 2016). Crops alternate between soybeans and corn on an annual basis, and tillage practices consist mainly of chisel plowing to a depth of 10 to 15 centimeters after corn harvest.

The agricultural field has been irrigated with treated domestic septage for more than 40 years. The septage, consisting almost entirely of residential wastewater from septic and holding tanks, is sprayed from a tanker truck onto the soil and row crops. Septage is applied year round, mostly during the growing season and fall before winter freezeup. An average of 2.25 million liters is applied every year across the field at a rate of about 160,000 liters per hectare. About 50 percent of the septage receives a lime stabilization treatment for 30 minutes to raise the pH of the wastewater to about 12 to control pathogens before it is applied to land. Biosolids have been applied to the field previously; however, this practice has not been used for at least two decades. Other sources of irrigation water have not been applied to the field.

Methods

The methods section describes field procedures used to collect soil samples, laboratory analytical methods, and laboratory quality-assurance and quality-control samples and analyses.

Soil Sample Collection Methods

Soil samples were collected from four soil sites around the perimeter of the agricultural field (fig. 1). Soil representative of the top 15 centimeters was scooped into a stainless steel bowl using a stainless steel scoop. Once enough soil was collected to fill a 500 milliliter jar, the sample was manually



Figure 1. Location of soil sites on an agricultural field sampled for wastewater indicators, hormones, sterols, antibiotics, and pharmaceuticals, central Minnesota, September 2014.

stirred with the stainless steel scoop to homogenize it before dispensing it into a glass amber jar. The stainless steel bowl and scoop were cleaned following USGS protocols for organic chemical sampling (Wilde, 2004). Briefly, the bowl and scoop were washed with, in succession, Liquinox $(\mathbb{R} + tap)$ water solution, tap water, methanol, and organic-blank water obtained from the USGS National Water Quality Laboratory in Lakewood, Colorado, prior to initial sample collection and in between collection of soil from different sites. Soil samples were frozen at -4 degrees Celsius until shipped to the analyzing laboratories.

Laboratory Analytical Methods

Soil samples were analyzed for 60 wastewater indicators (including 3 surrogate standards) and 33 hormones and sterols (including 13 surrogate standards) at the U.S. Geological

Survey National Water Quality Laboratory in Lakewood, Colorado, and 37 antibiotics and pharmaceuticals at the U.S. Geological Survey Organic Geochemistry Research Laboratory in Lawrence, Kansas. Wastewater indicators were determined by gas chromatography/mass spectrometry (Burkhardt and others, 2006), and hormones and sterols were determined by gas chromatography/tandem mass spectrometry by an adaptation of the method of Foreman and others (2012) for solid samples (Yang and others, 2012). Antibiotics and pharmaceuticals were determined by a research method previously described by Gibs and others (2013), Massey and others (2010), McKinney and others (2010), and Watanabe and others (2010). Briefly, samples were thawed and freeze-dried to remove moisture. Antibiotics and pharmaceuticals were extracted from the freeze-dried samples with a citric acid buffer adjusted to pH 6.0 with sodium hydroxide and mixed 50/50 (volume/volume) with methanol. Extracts were evaporated to

about 50 percent of their original volume using a TurboVap LV nitrogen evaporator. Chemical determinations were made by using a liquid chromatography method with tandem mass spectrometry in electrospray positive and negative mode (for two compounds) with scheduled multiple reaction monitoring and negative multiple reaction monitoring modes. A complete list of antibiotics and pharmaceuticals included in analysis is provided in table 1. The reporting level for most antibiotics and pharmaceuticals was 1 microgram per kilogram (μ g/kg). Reporting levels for ibuprofen and virginiamycin were 50 and 5 μ g/kg, respectively.

Laboratory Quality-Assurance Methods and Results

Several quality-assurance measures were done with analyses including laboratory reagent-spike samples, laboratory-blank samples, one laboratory duplicate sample, and one matrix-spike sample in which a baked reagent-sand matrix was spiked with chemicals of interest. Additional quality assurance was provided by surrogates, isotope dilution standards, or both, which were added to all samples before analysis. Relative percent difference between the laboratory duplicate samples was calculated as:

$$RPD = (|C_1 - C_2| / [(C_1 + C_2)/2]) \times 100$$
(1)

where

RPD

is the relative percent difference;

- C_1 is the measured concentration in sample 1, in micrograms per kilogram; and
- C_2 is the measured concentration in sample 2, in micrograms per kilogram.

Similarly, the percent recovery of the laboratory baked reagent-sand matrix-spike sample was calculated as:

$$PR = (C_{sp} - C_{env}/C_{exp}) \times 100$$
⁽²⁾

where

PR is the percent recovery,

- C_{sp} is the measured concentration, in micrograms per kilogram, in the spiked environmental sample,
- C_{env} is the measured concentration, in micrograms per kilogram, in the unspiked environmental sample, and
- C_{exp} is the nominal concentration, in micrograms per kilogram, added to the unspiked environmental sample.

Average percent recovery of 19 wastewater indicators was below 60 percent. Additionally, 4-nonylphenol, 4-nonylphenol diethoxylate, and 4-nonylphenol monoethoxylate had a percent recovery of zero. Additionally, average percent recovery of the antibiotics doxycycline and ofloxacin was 50 percent. Measurable concentrations of 10 analytes were detected in laboratory-blank samples; most maximum concentrations ranged from 3 to 9 times less than the respective reporting levels. However, the maximum concentration of phenol, 78.2 µg/kg, detected in laboratory-blank samples was above the reporting level of 50 μ g/kg. All environmental detections of phenol in soil samples were coded with a "v" to reflect elevated concentrations found in laboratory blanks and were not included in counts of detections. The percent recovery of all analytes in the laboratory matrix-spike sample varied greatly, ranging from 1.1 to 418 percent. Average laboratory matrix-spike percent recoveries for antibiotics and pharmaceuticals were less than wastewater indicator chemicals, which were less than hormones, sterols, and bisphenol A. Average percent recovery of surrogate or isotope dilution standards was often below 60 percent, suggesting that reported concentrations not corrected for isotope dilution standards potentially underestimated environmental concentrations in the samples.

Presence of Wastewater Indicators, Hormones, Sterols, Antibiotics, and Pharmaceuticals in Soil Irrigated with Domestic Septage

A total of 32 contaminants were detected among the 4 soil samples (table 2). Quantifiable concentrations ranged from 0.21 to about 2,460 µg/kg; concentrations of the antibiotics ciprofloxacin and ofloxacin were greater than the limit of quantification $(1,000 \ \mu g/kg)$ for soil sites 2 and 3. Detectable concentrations of hormones, sterols, and wastewater indicators generally were comparable to ranges reported in bottom sediments collected from Minnesota streams and rivers downstream from wastewater treatment plant effluent discharge (fig. 2; Lee and others 2011; Elliott and others 2016). Of the detected contaminants, 56 percent were detected in at least two samples, and 37 percent were detected in all four. A greater percentage of hormones and sterols were detected (47 percent) compared to wastewater indicators (29 percent) or antibiotics and pharmaceuticals (19 percent). Given the high soil adsorption coefficients for 4-androstene-3,17-dione, cis-androsterone, and 3β -coprostanol, these chemicals are more likely to adsorb to soils and have lower mobility to deeper depths or groundwater and may help to explain their presence in all the soil samples.

Of the hormones analyzed, three androgens (4-androstene-3,17-dione, *cis*-androsterone, and dihydrotestosterone) and two estrogens (estrone and progesterone) were detected. Total androgen and estrogen concentrations in each soil sample ranged from 1.5 to 10.6 and 0.2 to 0.9 μ g/kg, respectively. All detectable concentrations of *cis*-androsterone in this study were greater than the 75th percentile of those reported for river bottom sediments in Minnesota (fig. 2; Lee and others 2011; Elliott and others 2016). In addition to the detected biogenic estrogens, several other contaminants were **Table 1.** Antibiotics and pharmaceuticals analyzed in soil samples collected from a Central Minnesota agricultural field,September 2014.

[CASRN, Chemical Abstracts Service Registry Number; µg/kg, micrograms per kilogram; USGS, U.S. Geological Survey; OGRL, Organic Geochemistry Research Laboratory; --, no data]

| | Chemical | CASRN ¹ | Reporting level,² in µg/kg |
|---|----------|--------------------|-------------------------------|
| Antibiotics and pharmaceuticals analyzed at USGS OGRL | | | |
| Azithromycin | | 117772-70-0 | 1 |
| Carbamazepine | | 298-46-4 | 1 |
| Chloramphenicol | | 56-75-7 | 1 |
| Chlorotetracycline | | 64-72-2 | 1 |
| Ciprofloxacin | | 85721-33-1 | 1 |
| Doxycycline | | 564-25-0 | 1 |
| Enrofloxacin | | 93106-60-6 | 1 |
| Epichlorotetracycline | | | 1 |
| Epiisochlorotetracycline | | | 1 |
| Epioxytetracycline | | | 1 |
| Epitetracycline | | 79-85-6 | 1 |
| Erythromycin | | 114-07-8 | 1 |
| Erythromycin-H ₂ O | | 23893-13-2 | 1 |
| Ibuprofen | | 15687-27-1 | 50 |
| Isochlorotetracycline | | 514-53-4 | 1 |
| Lincomycin | | 154-21-2 | 1 |
| Lomefloxacin | | 98079-51-7 | 1 |
| Norfloxacin | | 70458-96-7 | 1 |
| Ofloxacin | | 82419-36-1 | 1 |
| Ormetoprim | | 6981-18-6 | 1 |
| Oxytetracycline | | 6153-64-6 | 1 |
| Roxithromycin | | 80214-83-1 | 1 |
| Sarafloxacin | | 98105-99-8 | 1 |
| Sulfachloropyridazine | | 80-32-0 | 1 |
| Sulfadiazine | | 68-35-9 | 1 |
| Sulfadimethoxine | | 122-11-2 | 1 |
| Sulfamethazine | | 57-68-1 | 1 |
| Sulfamethoxazole | | 723-46-6 | 1 |
| Sulfathiazole | | 72-14-0 | 1 |
| Tetracycline | | 60-54-8 | 1 |
| Total chlorotetracycline | | | 1 |
| Total erythromycin | | | 1 |
| Total oxytetracycline | | | 1 |
| Total tetracycline | | | 1 |
| Trimethoprim | | 738-70-5 | 1 |
| Tylosin | | 1401-69-0 | 1 |
| Virginiamycin | | 11006-76-1 | 5 |

¹This report contains Chemical Abstracts Service Registry Numbers (CASRN)®, which is a Registered Trademark of the American Chemical Society. The CASRN online database provides the latest registry number information: http://www.cas.org/. Chemical Abstracts Service recommends the verification of the CASRNs through Chemical Abstracts Service Client ServicesSM.

²U.S. Geological Survey Organic Geochemistry Research Laboratory.

Table 2. Concentrations of wastewater indicators, hormones, sterols, antibiotics, and pharmaceuticals in soil samples collected from a Minnesota agricultural field that applies domestic septage to the land, central Minnesota, September 2014.

[Table only includes chemicals that were detected in at least one sample. See figure 1 for soil site locations. na, not applicable; nr, not reported—sample potentially affected by laboratory contamination; <, less than; E, estimated; >, greater than]

| Chomical | Poporting lovel | Sample concentration, in micrograms per kilogram | | | |
|---|--------------------------------|--|-------------|-------------|-------------|
| | Keporting level ¹ – | Soil site 1 | Soil site 2 | Soil site 3 | Soil site 4 |
| | Was | stewater indicator | rs | | |
| Sample weight (grams) | na | 9.9 | 10 | 10 | 10 |
| <i>p</i> -Cresol | 250 | 50 | 40 | 40 | nr |
| 4-tert-Octylphenol | 50 | <50 | <50 | 6 | <50 |
| BDE congener 47 | 50 | <50 | <50 | 3 | <50 |
| Tributyl phosphate | 50 | <50 | 10 | <50 | <50 |
| 3-Methyl-1 <i>H</i> -indole | 50 | 7 | 10 | 5 | 4 |
| Acetyl hexamethyl tetrahydronaphthalene | 50 | <50 | <50 | 11 | 2 |
| Indole | 100 | 60 | 110 | 90 | 70 |
| Isophorone | 50 | <50 | <50 | 11 | 6 |
| Carbazole | 50 | <50 | 12 | 6 | 4 |
| 9,10-Anthraquinone | 50 | <50 | 10 | 7 | 5 |
| Acetophenone | 150 | nr | nr | 120 | nr |
| 2,6-Dimethylnaphthalene | 50 | <50 | <50 | <50 | 1 |
| Benzo[<i>a</i>]pyrene | 50 | 16 | 48 | 14 | 20 |
| Fluoranthene | 50 | 26 | 65 | <50 | <50 |
| Pyrene | 50 | 21 | 59 | 20 | 29 |
| Triclosan | 50 | <50 | nr | 17 | 6 |
| Bisphenol A ² | 50 | <50 | 13 | nr | nr |
| 3β-Coprostanol ² | 500 | 1,140 | 1,470 | nr | nr |
| β-Sitosterol | 500 | 1,550 | 1,300 | <500 | <500 |
| | Hormones | , sterols, and bisp | henol A | | |
| Sample weight (grams) | na | 4.99 | 4.83 | 4.97 | 4.98 |
| 4-Androstene-3,17-dione | 0.1 | 3.65 | 1.31 | 0.59 | < 0.43 |
| cis-Androsterone | 0.25 | 3.08 | 1.72 | 1.13 | 0.62 |
| Dihydrotestosterone | 0.1 | 3.9 | 1.82 | 1.13 | 0.9 |
| Estrone | 0.1 | < 0.36 | < 0.34 | < 0.20 | 0.21 |
| Progesterone | 0.5 | <1.05 | <1.45 | 0.91 | <1.01 |
| Bisphenol A ² | 10 | E 35 | E 27.7 | E 10.9 | E 36.3 |
| 3β-Coprostanol ² | 50 | E 1,850 | E 2,460 | E 1,655 | 1,352 |
| Cholesterol | 50 | 702 | E 1,031 | 731 | 326 |
| | Antibiotio | cs and pharmace | uticals | | |
| Sample weight (grams) | na | 1 | 1 | 1 | 1 |
| Carbamazepine | 1.0 | 1.4 | 4.3 | 6.3 | 3.9 |
| Norfloxacin | 1.0 | <1 | 10 | 24 | <1 |
| Ciprofloxacin | 1.0 | 380 | >1,000 | >1,000 | 540 |
| Ofloxacin | 1.0 | 400 | >1,000 | >1,000 | 310 |
| Epitetracycline | 1.0 | <1 | 3.1 | 5.7 | 1.8 |
| Tetracycline | 1.0 | <1 | 3.3 | 6.6 | 2.2 |
| Azithromycin | 1.0 | <1 | 180 | 540 | <1 |

¹Determined by the U.S. Geological Survey National Water Quality Laboratory or Organic Geochemistry Research Laboratory.

²Analyzed by two different analytical methods.



Figure 2. Comparison of concentrations for wastewater indicators, hormones, and sterols detected in a central Minnesota agricultural soil with bottom sediments collected from receiving streams, downstream from effluent inputs from wastewater treatment plants. Data for bottom sediments are from Elliott and others (2016) and Lee and others (2011).

detected that exhibit estrogenic properties: 4-*tert*-octylphenol, tributyl phosphate, acetyl hexamethyl tetrahydronaphthalene, and bisphenol A. Although these contaminants have relatively weak estrogenic properties by themselves, the additive estrogenicity of weak estrogens in a mixture may increase the effective estrogenicity of the soil (Silva and others, 2002).

Six antibiotics and one anticonvulsant were detected at least once among all soil samples. Carbamazepine, ciprofloxacin, and ofloxacin were detected in all soil samples; the sum of concentrations ranged from 201 to greater than 2,500 µg/kg per sample. Two fluoroquinolones (ciprofloxacin and oflaxacin) were present in all four soil samples at concentrations ranging from 310 to greater than 1,000 µg/kg. Azithromycin (antibiotic) was present in two soil samples at concentrations of 180 and 540 µg/kg. Ciprofloxacin and ofloxacin tend to have high sorption to soils, even soils with low organic carbon and high sand contents (Leal and others, 2013; Peng and others, 2014). Additionally, ofloxacin can inhibit microbial growth by more than 50 percent at concentrations of 5,000 μ g/kg (Peng and others, 2014). Carbamazepine is detected frequently in soils from fields irrigated with wastewater at concentrations similar to those observed in this study (Durán-Alvarez and others, 2009; Kinney and others, 2006). Compared to other studies, soil concentrations of tetracycline in this study are relatively low (2–6 compared to 3–20 μ g/kg; Chen and others, 2011).

Fluoroquinolones and other antibiotics persist in agricultural soils and earthworms amended with biosolids (which often contain the same contaminants as septage) with the potential for accumulation over time (Golet and others, 2003; Kinney and others, 2008). The presence of antibiotics in soils also may lead to shifts in microbial communities dominated by resistant organisms (Thiele-Bruhn, 2003) or favor particular bacteria, altering the microbial activity of soils (Córdova-Kreylos and Scow, 2007). Microbial abundance and diversity can shift when exposed to ciprofloxacin concentrations as low as 200 μ g/kg (Girardi and others, 2011). Ciprofloxacin concentrations in the soil in the current study were all greater than 200 μ g/kg; two were an order of magnitude greater, indicating the potential for an altered microbial community.

Hydraulic factors such as overlying soil properties and degree of aquifer confinement are important factors in the transport of organic contaminants to groundwater (Lapworth and others 2012); for example, carbamazepine is more likely to sorb to soils with high clay or organic matter content (Arye and others, 2011). The soil organic matter content was not analyzed, but the Natural Resources Conservation Service provides estimates of less than 2 percent organic matter at sites 1–3 and 6 percent at site 4 (Natural Resources Conservation Service, U.S. Department of Agriculture, 2017). Sorption to soils has been an important mechanism for removal of contaminants such as carbamazepine and sulfamethoxazole from water (Martínez-Hernández and others, 2016); in fact, carbamazepine was detected in all the soil samples (table 2).

Summary

Traditional on-site wastewater treatment systems are not designed to remove wastewater indicators, hormones, sterols, antibiotics, and pharmaceuticals, yet treated domestic septage can be used to irrigate agricultural fields. Soil collected from a central Minnesota agricultural field irrigated with domestic septage was analyzed for a variety of contaminants composed of 60 wastewater indicators (including 3 surrogate standards) and 33 hormones and sterols (including 13 surrogate standards) at the U.S. Geological Survey National Water Quality Laboratory in Lakewood, Colorado, and 37 antibiotics and pharmaceuticals at the U.S. Geological Survey Organic Geochemistry Research Laboratory in Lawrence, Kansas. A total of 32 contaminants were detected among the samples collected from four soil sites. In total, 19 wastewater indicators were detected at concentrations ranging from 1 (2,6-dimethylnaphthalene at soil site 4) to 1,550 (β -sitosterol at soil site 1) micrograms per kilogram (µg/kg). Considering the number of individual compounds analyzed within each chemical group (wastewater indicators; hormones, sterols, and bisphenol A; and antibiotics and pharmaceuticals), hormones and sterols were detected the most frequently (47 percent). Quantifiable concentrations of androgens and estrogens ranged from 0.21 (estrone at soil site 3) to 3.9 (dihydrotestosterone at soil site 1) μ g/kg, and androgens were more prevalent than estrogens. A total of seven antibiotics and pharmaceuticals were detected at concentrations ranging from 1.4 μ g/kg (carbamazepine at soil site 1) to above the limit of quantification, 1,000 μ g/kg (ciprofloxacin and ofloxacin at soil sites 2 and 3). Results from this pilot sampling indicate that contaminants may be accumulating in soil that is lacking a high organic content. Further research needs to be done to assess fate and transport of wastewater indicators, hormones, sterols, antibiotics, and pharmaceuticals through the soil column that may potentially reach groundwater, as well as effect soil health.

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Appendix 1

The appendix consists of one table listing wastewater indicators, hormones, sterols, antibiotics, and pharmaceuticals analyzed (table 1.1), and five tables summarizing laboratory quality-assurance data for reagent-spike samples (table 1.2), laboratory-blank samples (table 1.3), duplicate samples (table 1.4), matrix-spike samples (table 1.5), and surrogate and isotope dilution standards (table 1.6).

Table 1.1Wastewater indicators, hormones, and sterols analyzed in soil samples collected from an agricultural field, central
Minnesota, September 2014.

[CASRN, Chemical Abstracts Service Registry Number; µg/kg, micrograms per kilogram; USGS, U.S. Geological Survey; NWQL, National Water Quality Laboratory; --, no data; na, not applicable; d, deuterium]

| Chemical | CASRN ¹ | Reporting level,² in µg/kg |
|--|--------------------|-------------------------------|
| Wastewater indicators analyzed at USG | S NWQL | |
| 1,4-Dichlorobenzene | 106-46-7 | 100 |
| 1-Methylnaphthalene | 90-12-0 | 100 |
| 2,6-Dimethylnaphthalene | 581-42-0 | 100 |
| 2-Methylnaphthalene | 91-57-6 | 100 |
| 3β-Coprostanol | 360-68-9 | 1,000 |
| 3-Methyl-1 <i>H</i> -indole (Skatole) | 83-34-1 | 100 |
| 3-tert-Butyl-4-hydroxyanisole (BHA) | 121-00-6 | 300 |
| 4-Cumylphenol | 599-64-4 | 100 |
| 4-n-Octylphenol | 1806-26-4 | 100 |
| 4-Nonylphenol (sum of all isomers) | | 1,500 |
| 4-Nonylphenol diethoxylate (sum of all isomers; NP2EO) | 20427-84-3 | 2,000 |
| 4-Nonylphenol monoethoxylate (sum of all isomers; NP1EO) | 68412-54-4 | 1,000 |
| 4-tert-Octylphenol | 140-66-9 | 100 |
| 4-tert-Octylphenol diethoxylate (OP2EO) | 2315-61-9 | 100 |
| 4-tert-Octylphenol monoethoxylate, (OP1EO) | 2315-67-5 | 500 |
| 9,10 Anthraquinone | 84-65-1 | 100 |
| Acetophenone | 98-86-2 | 300 |
| Acetyl hexamethyl tetrahydro naphthalene (AHTN) | 21145-77-7 | 100 |
| Anthracene | 120-12-7 | 100 |
| Atrazine | 1912-24-9 | 200 |
| BDE congener 47 | 5436-43-1 | 100 |
| Benzo[a]pyrene | 50-32-8 | 100 |
| Benzophenone | 119-61-9 | 100 |
| β-Sitosterol | 83-46-5 | 1,000 |
| β-Stigmastanol | 19466-47-8 | 1,000 |
| Bis(2-ethylhexyl) phthalate | 117-81-7 | 500 |
| Bisphenol A | 80-05-7 | 100 |
| Bromacil | 314-40-9 | 1,000 |
| Camphor | 76-22-2 | 100 |
| Carbazole | 86-74-8 | 100 |
| Chlorpyrifos | 2921-88-2 | 100 |
| Cholesterol | 57-88-5 | 500 |
| Diazinon | 333-41-5 | 100 |
| Diethyl phthalate | 84-66-2 | 200 |
| D-Limonene | 5989-27-5 | 100 |
| Fluoranthene | 206-44-0 | 100 |
| Hexahydrohexamethyl cyclopentabenzopyran (HHCB) | 1222-05-5 | 100 |
| Indole | 120-72-9 | 200 |
| Isoborneol | 124-76-5 | 100 |
| Isophorone | 78-59-1 | 100 |

Table 1.1Wastewater indicators, hormones, and sterols analyzed in soil samples collected from an agricultural field, central
Minnesota, September 2014.—Continued

[CASRN, Chemical Abstracts Service Registry Number; μg/kg, micrograms per kilogram; USGS, U.S. Geological Survey; NWQL, National Water Quality Laboratory; --, no data; na, not applicable; d, deuterium]

| Chemical | CASRN ¹ | Reporting level,² in µg/kg |
|--|--------------------|-------------------------------|
| Wastewater indicators analyzed at USGS NWQL— | Continued | |
| Isopropylbenzene | 98-82-8 | 200 |
| Isoquinoline | 119-65-3 | 200 |
| Menthol | 89-78-1 | 100 |
| Metolachlor | 51218-45-2 | 100 |
| Naphthalene | 91-20-3 | 100 |
| N,N-Diethyl- <i>m</i> -toluamide (DEET) | 134-62-3 | 200 |
| <i>p</i> -Cresol | 106-44-5 | 500 |
| Phenanthrene | 85-01-8 | 100 |
| Phenol | 108-95-2 | 100 |
| Prometon | 1610-18-0 | 100 |
| Pyrene | 129-00-0 | 100 |
| Tributyl phosphate | 126-73-8 | 100 |
| Triclosan | 3380-34-5 | 100 |
| Triphenyl phosphate | 115-86-6 | 100 |
| Tris(2-butoxyethyl) phosphate | 78-51-3 | 300 |
| Tris(2-chloroethyl) phosphate | 115-96-8 | 200 |
| Tris(dichloroisopropyl) phosphate | 13674-87-8 | 200 |
| Bisphenol A- d_{14} (surrogate) | | na |
| Decafluorobiphenyl (surrogate) | 434-90-2 | na |
| Fluoranthene- d_{10} (surrogate) | 93951-69-0 | na |
| Steroid hormones, sterols, and bisphenol A analyzed at | USGS NWQL | |
| 11-Ketotestosterone | 564-35-2 | 0.52 |
| 17α-Estradiol | 57-91-0 | 0.2 |
| 17α-Ethynylestradiol | 57-63-6 | 0.2 |
| 17β-Estradiol | 50-28-2 | 0.4 |
| 3β-Coprostanol | 360-68-9 | 50 |
| 4-Androstene-3,17-dione | 63-05-8 | 0.5 |
| Bisphenol A | 80-05-7 | 20 |
| Cholesterol | 57-88-5 | 120 |
| cis-Androsterone | 53-41-8 | 0.5 |
| Dihydrotestosterone | 521-18-6 | 1 |
| Epitestosterone | 481-30-1 | 1 |
| Equilenin | 517-09-9 | 0.52 |
| Equilin | 474-86-2 | 4 |
| Estriol | 50-27-1 | 0.52 |
| Estrone | 53-16-7 | 0.5 |
| Mestranol | 72-33-3 | 0.4 |
| Norethindrone | 68-22-4 | 0.4 |
| Progesterone | 57-83-0 | 3 |
| Testosterone | 58-22-0 | 0.4 |

Table 1.1Wastewater indicators, hormones, and sterols analyzed in soil samples collected from an agricultural field, centralMinnesota, September 2014.—Continued

[CASRN, Chemical Abstracts Service Registry Number; µg/kg, micrograms per kilogram; USGS, U.S. Geological Survey; NWQL, National Water Quality Laboratory; --, no data; na, not applicable; d, deuterium]

| Chemical | CASRN ¹ | Reporting level,² in µg/kg | | | |
|--|--------------------|-------------------------------|--|--|--|
| Steroid hormones, sterols, and bisphenol A analyzed at USGS NWQL—Continued | | | | | |
| trans-Diethylstilbestrol | 56-53-1 | 0.33 | | | |
| 16-Epiestriol-2, $4-d_2$ (surrogate) | 366495-94-5 | na | | | |
| 17α -Ethynylestradiol-2, 4, 16, 16- d_4 (surrogate) | 350820-06-3 | na | | | |
| 17β-Estradiol-13,14,15,16,17,18- ¹³ C ₆ (surrogate) | | na | | | |
| Bisphenol A- d_{16} (surrogate) | 96210-87-6 | na | | | |
| Cholesterol-25, 26, 26, 26, 27, 27, 27- d_{γ} (surrogate) | 83199-47-7 | na | | | |
| <i>cis</i> -Androsterone-16,16- d_2 (surrogate) | 89685-22-3 | na | | | |
| Estriol-2, 4, 16, 17- d_4 (surrogate) | | na | | | |
| Estrone-13,14,15,16,17,18- ¹³ C ₆ (surrogate) | | na | | | |
| Medroxyprogesterone- d_3 (surrogate) | 162462-69-3 | na | | | |
| Mestranol-2, 4, 16, 16- d_4 (surrogate) | | na | | | |
| Nandrolone-16,16,17-d ₃ (surrogate) | 120813-22-1 | na | | | |
| Progesterone-2, 3, $4^{-13}C_3$ (surrogate) | 327048-87-3 | na | | | |
| <i>trans</i> -Diethy <i>l</i> -1,1,1',1'-d ₄ -stilbesterol-3,3',5,5'-d ₄ (surrogate) | | na | | | |

¹This report contains Chemical Abstracts Service Registry Numbers (CASRN)®, which is a Registered Trademark of the American Chemical Society. The CASRN online database provides the latest registry number information: http://www.cas.org/. Chemical Abstracts Service recommends the verification of the CASRNs through Chemical Abstracts Service Client ServicesSM.

²Determined by the U.S. Geological Survey National Water Quality Laboratory.

Table 1.2Recovery of method analytes in laboratory reagent-spike samples analyzed at the U.S. Geological Survey National WaterQuality Laboratory and Organic Geochemistry Research Laboratory.

[CASRN, Chemical Abstracts Service Registry Number; USGS, U.S. Geological Survey; NWQL, National Water Quality Laboratory; m, highly variable chemical using this method, questionable precision and (or) accuracy; E, estimated value; v, chemical detected in laboratory blank; --, no data; Organic Geochemistry Research Laboratory]

| | | Percen | Percent recovery | | |
|---|---------------------------|---------------------------|-------------------------------|--|--|
| Chemical | CASRN ¹ | Analyzed March 3, 2015 | Analyzed February 19, 2015 | | |
| Wastewater indicator comp | ounds analyzed at USGS NW | OL | | | |
| 1,4-Dichlorobenzene | 106-46-7 | 61.6 m | 64.9 m | | |
| 1-Methylnaphthalene | 90-12-0 | 69 | 89 | | |
| 2,6-Dimethylnaphthalene | 581-42-0 | 73 | 93 | | |
| 2-Methylnaphthalene | 91-57-6 | 67 | 90 | | |
| 3β-Coprostanol | 360-68-9 | 0 m | E105 vm | | |
| 3-Methyl-1 <i>H</i> -indole (Skatole) | 83-34-1 | 83 | 98 | | |
| 3-tert-Butyl-4-hydroxy anisole (BHA) | 121-00-6 | 19 m | 12 m | | |
| 4-Cumylphenol | 599-64-4 | 105 | 117 | | |
| 4-n-Octylphenol | 1806-26-4 | 94 | 90 | | |
| 4-Nonylphenol (sum of all isomers) | | 0 m | 0 m | | |
| 4-Nonylphenol diethoxylate (NP2EO, all isomers) | 20427-84-3 | 0 m | 0 m | | |
| 4-Nonylphenol monoethoxylate (NP1EO, all isomers) | 68412-54-4 | 0 m | E154 vm | | |
| 4-tert-Octylphenol | 140-66-9 | 87 | 102 | | |
| 4-tert-Octylphenol diethoxylate (OP2EO) | 2315-61-9 | E126 vm | E119 m | | |
| 4-tert-Octylphenol monoethoxylate (OP1EO) | 2315-67-5 | E60 vm | E64 vm | | |
| Acetophenone | 98-86-2 | E63 vm | 108 m | | |
| Acetyl hexamethyl tetrahydronaphthalene (AHTN) | 21145-77-7 | 108 | 118 | | |
| Anthracene | 120-12-7 | E91 vm | 102 | | |
| Anthraquinone | 84-65-1 | 41 | 84 | | |
| Atrazine | 1912-24-9 | 62 | 126 | | |
| BDE congener 47 | 5436-43-1 | 108 m | 111 m | | |
| Benzo[a]pyrene | 50-32-8 | 97 | 112 | | |
| Benzophenone | 119-61-9 | E91 v | E112 v | | |
| β-Sitosterol | 83-46-5 | 0 m | 0 m | | |
| β-Stigmastanol | 19466-47-8 | 0 m | 0 m | | |
| Bis(2-ethylhexyl) phthalate | 117-81-7 | E126 v | E104 v | | |
| Bisphenol A | 80-05-7 | E74 m | 62 m | | |
| Bromacil | 314-40-9 | 48 m | 105 m | | |
| Camphor | 76-22-2 | 51.4 | 101 | | |
| Carbazole | 86-74-8 | 100 | 119 | | |
| Chlorpyrifos | 2921-88-2 | 62 m | 52 m | | |
| Cholesterol | 57-88-5 | 0 m | 0 m | | |
| Diazinon | 333-41-5 | 2 | 2 | | |
| Diethyl phthalate | 84-66-2 | 0 | E58 v | | |
| D-Limonene | 5989-27-5 | E53 vm | 60 m | | |
| Fluoranthene | 206-44-0 | E103 v | 114 | | |
| Hexahydrohexamethyl cyclopentabenzopyran (HHCB) | 1222-05-5 | 109 | 113 | | |
| Indole | 120-72-9 | 62 | 64 | | |

 Table 1.2
 Recovery of method analytes in laboratory reagent-spike samples analyzed at the U.S. Geological Survey National Water

 Quality Laboratory and Organic Geochemistry Research Laboratory.—Continued

[CASRN, Chemical Abstracts Service Registry Number; USGS, U.S. Geological Survey; NWQL, National Water Quality Laboratory; m, highly variable chemical using this method, questionable precision and (or) accuracy; E, estimated value; v, chemical detected in laboratory blank; --, no data; Organic Geochemistry Research Laboratory]

| Chemical | | Percent recovery | | |
|-----------------------------------|---------------------------------|---------------------------|-------------------------------|--|
| | CASRN ¹ | Analyzed March 3, 2015 | Analyzed February 19, 2015 | |
| Wastewater indicator compo | unds analyzed at USGS NWQL—C | ontinued | | |
| Isoborneol | 124-76-5 | 0 m | 0 m | |
| Isophorone | 78-59-1 | 36 m | 75 m | |
| Isopropylbenzene | 98-82-8 | 0 m | 0 m | |
| Isoquinoline | 119-65-3 | 23 m | 86 m | |
| Menthol | 89-78-1 | 54 m | 108 m | |
| Metolachlor | 51218-45-2 | 92 | 119 | |
| Naphthalene | 91-20-3 | 75 | 93 | |
| N,N-Diethyl-m-toluamide (DEET) | 134-62-3 | 56 m | 118 m | |
| <i>p</i> -Cresol | 106-44-5 | 59 | 113 | |
| Phenanthrene | 85-01-8 | 96 | 105 | |
| Phenol | 108-95-2 | 0 m | | |
| Prometon | 1610-18-0 | 0 m | 104 m | |
| Pyrene | 129-00-0 | 102 | 112 | |
| Tributyl phosphate | 126-73-8 | E79 | 117 | |
| Triclosan | 3380-34-5 | 90 | 120 | |
| Triphenyl phosphate | 115-86-6 | 41 m | 22 m | |
| Tris(2-butoxyethyl) phosphate | 78-51-3 | 17 m | 120 m | |
| Tris(2-chloroethyl) phosphate | 115-96-58 | 43 m | 65 m | |
| Tris(dichloroisopropyl) phosphate | 13674-87-8 | 42 m | 17 m | |
| Hormones, sterols, and | bisphenol A analyzed at USGS NV | /QL | | |
| 11-Ketotestosterone | 564-35-2 | 91 | 106 | |
| 17α-Estradiol | 57-91-0 | 111 | 110 | |
| 17α-Ethynylestradiol | 57-63-6 | 100 | 104 | |
| 17β-Estradiol | 50-28-2 | 107 | 104 | |
| 3β-Coprostanol | 360-68-9 | 107 | 104 | |
| 4-Androstene-3,17-dione | 63-05-8 | 113 | 109 | |
| Bisphenol A | 80-05-7 | E135 | E130 | |
| Cholesterol | 57-88-5 | 113 | 109 | |
| <i>cis</i> -Androsterone | 53-41-8 | 101 | 98 | |
| Dihydrotestosterone | 521-18-6 | 108 | 108 | |
| Epitestosterone | 481-30-1 | 116 | 113 | |
| Equilenin | 517-09-9 | 114 | 104 | |
| Equilin | 474-86-2 | 100 m | 88 m | |
| Estriol | 50-27-1 | E95 | 109 | |
| Estrone | 53-16-7 | 113 | 107 | |
| Mestranol | 72-33-3 | 99 | 105 | |
| Norethindrone | 68-22-4 | 116 | 98 | |

Table 1.2 Recovery of method analytes in laboratory reagent-spike samples analyzed at the U.S. Geological Survey National Water

 Quality Laboratory and Organic Geochemistry Research Laboratory.—Continued

[CASRN, Chemical Abstracts Service Registry Number; USGS, U.S. Geological Survey; NWQL, National Water Quality Laboratory; m, highly variable chemical using this method, questionable precision and (or) accuracy; E, estimated value; v, chemical detected in laboratory blank; --, no data; Organic Geochemistry Research Laboratory]

| Chemical | | Percent recovery | |
|-------------------------------|---|---------------------------|-------------------------------|
| | CASRN ¹ | Analyzed March 3, 2015 | Analyzed February 19, 2015 |
| Hormones, sterols, a | nd bisphenol A analyzed at USGS NWQL—Co | ntinued | |
| Progesterone | 57-83-0 | 105 m | 97 m |
| Testosterone | 58-22-0 | 108 | 101 |
| trans-Diethylstilbestrol | 58-63-1 | E123 | E111 |
| Antibiotics a | and pharmaceuticals analyzed at USGS OGRL | | |
| Azithromycin | 117772-70-0 | 100 | |
| Carbamazepine | 298-46-4 | 96 | |
| Chloramphenicol | 56-75-7 | 82 | |
| Ciproflaxacin | 85721-33-1 | 69 | |
| Doxycycline | 564-25-0 | 50 | |
| Enrofloxacin | 93106-60-6 | 60 | |
| Erythromycin | 114-07-8 | 67 | |
| Erythromycin-H ₂ O | 23893-13-2 | 83 | |
| Ibuprofen | 15687-27-1 | 89 | |
| Lincomycin | 154-21-2 | 87 | |
| Lomefloxacin | 82419-36-1 | 120 | |
| Norfloxacin | 70458-96-7 | 69 | |
| Ofloxacin | 82419-36-1 | 50 | |
| Ormetoprim | 6981-18-6 | 98 | |
| Roxithromycin | 80214-83-1 | 91 | |
| Sarafloxacin | 98105-99-8 | 130 | |
| Sulfachloropyridazine | 80-32-0 | 82 | |
| Sulfadiazine | 68-35-9 | 100 | |
| Sulfadimethoxine | 122-11-2 | 70 | |
| Sulfamethazine | 57-68-1 | 91 | |
| Sulfamethoxazole | 723-46-6 | 90 | |
| Sulfathiazole | 72-14-0 | 87 | |
| Total chlorotetracycline | | 81 | |
| Total oxytetracycline | | 110 | |
| Total tetracycline | | 100 | |
| Trimethoprim | 738-70-5 | 100 | |
| Tylosin | 1401-69-0 | 85 | |
| Virginiamycin | 11006-76-1 | 82 | |

¹This report contains Chemical Abstracts Service Registry Numbers (CASRN)®, which is a Registered Trademark of the American Chemical Society. The CASRN online database provides the latest registry number information: http://www.cas.org/. Chemical Abstracts Service recommends the verification of the CASRNs through Chemical Abstracts Service Client ServicesSM.

Table 1.3Concentrations of wastewater indicators, hormones,
sterols, antibiotics, and pharmaceuticals in laboratory-blank soil
samples analyzed at the U.S. Geological Survey National Water
Quality Laboratory.

[Table only includes chemicals that were detected in blank samples; µg/kg, micrograms per kilogram]

| Chemical | Reporting level, in µg/kg | Maximum concentration, in µg/kg |
|---------------------------------|---------------------------------|---------------------------------------|
| 4-tert-Octylphenol diethoxylate | | |
| (OP2EO) | 100 | 14.6 |
| Anthracene | 100 | 10.5 |
| Benzophenone | 100 | 5.5 |
| β-Sitosterol | 1,000 | 79 |
| β-Stigmastanol | 1,000 | 51 |
| Cholesterol | 500 | 56 |
| D-Limonene | 100 | 5.9 |
| Fluoranthene | 100 | 12 |
| Phenanthrene | 100 | 9.2 |
| Phenol | 100 | 78.2 |

Table 1.4Relative percent difference between environmental(U.S. Geological Survey station 451700093430001, sampledSeptember 8, 2014) and laboratory duplicate samples analyzed atthe U.S. Geological Survey Organic Geochemistry Laboratory.

| Chemical | Relative percent difference |
|--------------------|--------------------------------|
| Azithromycin | 0 |
| Carbamazepine | 3 |
| Ciproflaxacin | 20 |
| Norfloxacin | 0 |
| Ofloxacin | 0 |
| Total tetracycline | 16 |

Table 1.5Percent recovery of study analytes in a laboratory matrix-spike soil sample (U.S. Geological Survey station 451700093430001,
sampled September 8, 2014).

[USGS, U.S. Geological Survey; OGRL, Organic Geochemistry Research Laboratory; --, no data—not included in spiked sample; NWQL, National Water Quality Laboratory]

| Chemical | Percent recovery | | |
|---|------------------|--|--|
| Antibiotics and pharmaceuticals analyzed at the USGS OGRL | | | |
| Azithromycin | 3.2 | | |
| Carbamazepine | 93 | | |
| Chloramphenicol | 100 | | |
| Ciproflaxacin | 1.8 | | |
| Doxycycline | 73 | | |
| Enrofloxacin | 1.1 | | |
| Ibuprofen | 100 | | |
| Lincomycin | 88 | | |
| Lomefloxacin | 15 | | |
| Norfloxacin | 2.0 | | |
| Ofloxacin | 2.3 | | |
| Ormetoprim | 87 | | |
| Roxithromycin | 66 | | |
| Sarafloxacin | 16 | | |
| Sulfachloropyridazine | 89 | | |
| Sulfadiazine | 86 | | |
| Sulfadimethoxine | 90 | | |
| Sulfamethazine | 94 | | |
| Sulfamethoxazole | 94 | | |
| Sulfathiazole | 86 | | |
| Total chlorotetracycline | 94 | | |
| Total erythromycin | 44 | | |
| Total oxytetracycline | 45 | | |
| Total tetracycline | 71 | | |
| Trimethoprim | 87 | | |
| Tylosin | 66 | | |
| Virginiamycin | | | |
| Wastewater indicator compounds analyzed at the USGS NWQL | | | |
| 1,4-Dichlorobenzene | 56 | | |
| 1-Methylnaphthalene | 92 | | |
| 2,6-Dimethylnaphthalene | 95 | | |
| 2-Methylnaphthalene | 96 | | |
| 3-Methyl-1 <i>H</i> -indole (Skatole) | 26 | | |
| 3β-Coprostanol | 80 | | |
| 3-tert-Butyl-4-hydroxy anisole (BHA) | | | |
| 4-Cumylphenol | 87 | | |
| 4-n-Octylphenol | 65 | | |
| 4-Nonylphenol (sum of all isomers) | 114 | | |
| 4-Nonylphenol diethoxylate (NP2EO, all isomers) | 161 | | |
| 4-Nonylphenol monoethoxylate (NP1EO, all isomers) | 190 | | |

 Table 1.5
 Percent recovery of study analytes in a laboratory matrix-spike soil sample (U.S. Geological Survey station 451700093430001, sampled September 8, 2014).—Continued

[USGS, U.S. Geological Survey; OGRL, Organic Geochemistry Research Laboratory; --, no data—not included in spiked sample; NWQL, National Water Quality Laboratory]

| Chemical | Percent recovery |
|--|------------------|
| Wastewater indicator compounds analyzed at the USGS NWC | DL—Continued |
| 4- <i>tert</i> -Octylphenol | 255 |
| 4- <i>tert</i> -Octylphenol diethoxylate (OP2EO) | 67 |
| 4- <i>tert</i> -Octylphenol monoethoxylate (OP1EO) | 70 |
| Acetophenone | 37 |
| Acetyl hexamethyl tetrahydronaphthalene (AHTN) | 95 |
| Anthracene | 130 |
| Anthraquinone | 139 |
| Atrazine | 109 |
| BDE congener 47 | 74 |
| Benzo[<i>a</i>]pyrene | 132 |
| Benzophenone | 137 |
| β-Sitosterol | 42 |
| β-Stigmastanol | 61 |
| Bis(2-ethylhexyl) phthalate | 87 |
| Bisphenol A | 62 |
| Bromacil | 69 |
| Camphor | 83 |
| Carbazole | 137 |
| Chlorpyrifos | 59 |
| Cholesterol | 122 |
| Diazinon | |
| Diethyl phthalate | 87 |
| D-Limonene | 25 |
| Fluoranthene | 245 |
| Hexahydrohexamethyl cyclopentabenzopyran (HHCB) | 80 |
| Indole | 75 |
| Isoborneol | |
| Isophorone | 25 |
| Isopropylbenzene | |
| Isoquinoline | 50 |
| Menthol | 79 |
| Metolachlor | 130 |
| Naphthalene | 100 |
| <i>N</i> , <i>N</i> -Diethyl- <i>m</i> -toluamide (DEET) | 50 |
| <i>p</i> -Cresol | 12 |
| Phenanthrene | 180 |
| Prometon | 89 |
| Pyrene | 220 |
| Tributyl phosphate | 145 |
| Triclosan | 122 |

 Table 1.5
 Percent recovery of study analytes in a laboratory matrix-spike soil sample (U.S. Geological Survey station 451700093430001, sampled September 8, 2014).—Continued

[USGS, U.S. Geological Survey; OGRL, Organic Geochemistry Research Laboratory; --, no data—not included in spiked sample; NWQL, National Water Quality Laboratory]

| Chemical | Percent recovery |
|------------------------------------|---------------------------------|
| Wastewater indicator compounds ana | yzed at the USGS NWQL—Continued |
| Triphenyl phosphate | 66 |
| Tris(2-butoxyethyl) phosphate | 130 |
| Tris(2-chloroethyl) phosphate | 31 |
| Tris(dichloroisopropyl) phosphate | 62 |
| Hormones, sterols, and bispheno | I A analyzed at the USGS NWQL |
| 11-Ketotestosterone | 213 |
| 17α-Estradiol | |
| 17α-Ethynylestradiol | 116 |
| 17β-Estradiol | |
| 3β-Coprostanol | 116 |
| 4-Androstene-3,17-dione | 418 |
| Bisphenol A | 103 |
| Cholesterol | 124 |
| cis-Androsterone | 98 |
| Dihydrotestosterone | 136 |
| Epitestosterone | 207 |
| Equilenin | |
| Equilin | 60 |
| Estriol | |
| Estrone | 141 |
| Mestranol | 102 |
| Norethindrone | 249 |
| Progesterone | 107 |
| Testosterone | 158 |

Table 1.6Percent recovery of surrogate and isotope dilution standards analyzed at the U.S. Geological Survey National Water Quality
Laboratory.

| | Percent recovery at U.S. Geological Survey station, sampled September 8, 2014 | | | | |
|---|---|---------------------------|-----------------|-----------------|--|
| Standard | 451700093430001 | 451700093440001 | 451700093430003 | 451700093430002 | |
| Wastewater indicator surrogate standards | | | | | |
| Bisphenol A- d_{14} | 10 | 20 | 20 | 40 | |
| Decafluorobiphenyl | 24 | 17 | 36 | 25 | |
| Fluoranthene- <i>d</i> ₁₀ | 62 | 65 | 83 | 84 | |
| Steroid hormor | ie, sterol, and bisphen | ol A isotope dilution sta | indards | | |
| 16-Epestriol-2,4-d ₂ | 6 | 19 | 36 | 29 | |
| 17α -Ethynylestradiol-2, 4, 16, 16- d_4 | 6 | 28 | 48 | 43 | |
| 17β-Estradiol-13,14,15,16,17,18- $^{13}C_6$ | 3 | 18 | 34 | 36 | |
| Bisphenol A- d_{16} | 5 | 18 | 40 | 28 | |
| Cholesterol-25,26,26,26,27,27,27-d ₇ | 33 | 23 | 26 | 59 | |
| <i>cis</i> -Androsterone-16,16-d ₂ | 94 | 43 | 57 | 18 | |
| Diethyl-1, 1, 1', 1'-d ₄ -stilbestrol-3, 3', 5, 5'-d ₄ | 1 | 3 | 5 | 2 | |
| Estriol-2,4,16,17-d ₄ | 3 | 10 | 10 | 3 | |
| Estrone-13, 14, 15, 16, 17, 18-13C ₆ | 8 | 35 | 53 | 48 | |
| Medroxyprogesterone- d_3 | 27 | 59 | 88 | 94 | |
| Mestranol-2, 4, 16, 16-d ₄ | 47 | 54 | 67 | 77 | |
| Nandrolone-16,16,17-d ₃ | 19 | 40 | 66 | 87 | |
| Progesterone-2, 3, $4^{-13}C_3$ | 16 | 42 | 58 | 53 | |
| <i>trans</i> -Diethyl-1,1,1'1'-d ₄ -stilbestrol-3,3',5,5'-d ₄ | 2 | 3 | 5 | 1 | |

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